

# Influence of natural regeneration on fractal features of residue microaggregates in bauxite residue disposal areas

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## INTRODUCTION

Residues from mineral ore processing are disposed on land in large residue disposal areas which may eventually create a series of ecological and environmental issues (Smart *et al.*, 2016; Wu *et al.*, 2016). In the aluminum industry, bauxite residue is an alkaline solid by-product generated when alumina is extracted from bauxite ore by the Bayer process (Goloran *et al.*, 2016; Kong *et al.*, 2017a). The global inventory has reached 3.4 billion tons, with an annual increase of 120 million tons (Xue *et al.*, 2016a; Kong *et al.*, 2017b). Large volumes of bauxite residue are deposited in bauxite residue disposal areas which cause potential environmental risks, as these bare areas are sensitive to erosion by wind and water, and can be regarded as a potential source of contamination due to their high alkalinity and salinity (Gelencsér *et al.* 2011; Ruyters *et al.* 2011). In situ rehabilitation and revegetation may however stabilize the residue surface and minimize wind erosion (Courtney *et al.*, 2009; Kaur *et al.*, 2016; Schmalenberger *et al.*, 2013). Its poor physical structure is nevertheless a major limitation to support plant growth (Liu *et al.*, 2013; Zhu *et al.*, 2016a). Residue particle sizes range from 2 to 2000  $\mu\text{m}$  and 60-80 % exist as  $<20 \mu\text{m}$  (Xue *et al.*, 2016b). Jones *et al.* (2011) reported that addition of organic waste may influence aggregate size distribution and increase the proportion of water-stable aggregates. Zhu *et al.* (2016b) found that natural regeneration may improve the physical condition and aggregate stability of bauxite residue.

Soil aggregate stability is one of the most important properties in soils and affects water erosion, soil aeration, nutrient recycling and biological activity, as well as plant growth (Cerdà 2000; Le Guillou *et al.*, 2012; Moncada *et al.*, 2013). Physical forces, chemical bonds and biological agents may drive the aggregation processes of soil particles (Yagüe *et al.*, 2016; Lehmann & Rillig, 2015). Microaggregate stability is an important soil property which is usually used to determine soil erosion resistance (Wang *et al.*, 2016). A better understanding of microaggregate formation is essential to maintain structural stability in soils. Several major binding agents such as clay minerals, organic carbon and polyvalent ions have significant effects on colloid flocculation (Zhou *et al.*, 2005). Barbosa *et al.* (2015) observed a cementation effect by organic carbon from poultry manure applications and clay flocculation enhancing aggregation. Igwe *et al.* (2009) discovered that oxalate and pyrophosphate extractable iron-aluminum oxides may act as aggregation agents to colloidal stability; organic carbon had acted in association with the oxides as a linkage with clay particles and polyvalent cations to enhance aggregate stability. Virto *et al.* (2008) concluded that stable microaggregates were formed within the silt-size fraction and organic carbon was stored by adsorption and entrapment of fine organic residues.

Soil structure is related to the size, shape and stability of soil aggregates (Aksakal *et al.*, 2016; Ahmadi *et al.*, 2011). Microaggregate stability depends on the size distribution of microaggregates and several procedures have been proposed for characterizing aggregate size distribution. Fractal theory is mainly used to analyze the relationship between local and overall irregular broken complex images and structural geometry under different scales. The concept of fractal dimension was proposed to provide a quantitative description for irregular shapes (Jing *et al.*, 2016). Fractal theory has been widely applied in soil science to quantify and estimate aggregate size distribution of soils (Kolay & Kayabali, 2006). Fractal dimension reveals the difference between particle size distribution and related physical properties (Wei *et al.*, 2016; Wang *et al.*, 2015). Gao *et al.* (2014) suggested that fractal dimension could be regarded as a considerable and reliable parameter to reflect variations in soil properties. High

1 values represent aggregates dominated by fine fragments, whilst low values represent large fragments.  
2 Many researchers have used fractal dimension to predict soil particle size distribution or the size  
3 distribution of water-stable aggregates (Peng *et al.*, 2014).

4 With the development of soil fractal theory, the limitation of single-fractal dimension has been  
5 stressed to describe soil particle size distribution. In order to obtain more detailed information of soil  
6 structure, multi-fractal theory was introduced to soil science (Li *et al.*, 2016). Rodríguez-Lado & Lado  
7 (2016) found that particle size distribution behaved as multi-fractals, with scaling properties varying in  
8 different soil samples, whilst values of fractal dimension may be related to the degree of evolution of  
9 the soils. Peng *et al.* (2014) found that the single- and multi-fractal parameters could describe soil  
10 particle size distribution and the influences of soil structure effectively. There are, however, few studies  
11 focusing on multi-fractal parameters of microaggregate size distribution.

12 This work focuses on an alumina refinery in Central China. The inventory of bauxite residue is an  
13 estimated 35 million tons, which is currently increasing by approximately 2.2 million tons per annum  
14 (Zhu *et al.*, 2016b). Bauxite ore is discharged in hot NaOH by the Bayer processes and the residues are  
15 pumped to the disposal areas using the dry stacking method. Spontaneous vegetation colonization over  
16 the past 20 years at the study site may reveal that natural weathering processes ameliorate the residue  
17 substrate and support plant growth. Natural regeneration also enhances the proportion of water-stable  
18 aggregates and resistance to erosion (Zhu *et al.*, 2016c). As microaggregate stability is used to predict  
19 soil surface erosion (Wang *et al.*, 2016), this study focus on 1) the effect of natural weathering  
20 processes on microaggregate stability of bauxite residue; 2) to evaluate microaggregate size  
21 distribution in bauxite residue by fractal parameters; 3) to investigate whether fractal parameters may  
22 be used as an indicator to evaluate microaggregate stability of bauxite residue.

## 23 MATERIALS AND METHODS

### 24 *Soil Sampling*

25 Residue samples were collected from a disposal area in Central China. The climate is temperate  
26 continental monsoon, with a mean annual daily temperature of 12.8°C-14.8°C and average precipitation  
27 ranging from 600 mm to 1200 mm per year.

28 According to ecological field investigations, five different zones related to disposal age were  
29 selected during August to September 2014. These included (a) 1-year-old bauxite residue (R1), (b) 4-  
30 year-old bauxite residue (R2), (c) 6-year-old bauxite residue (R3), (d) 10-year-old bauxite residue (R4),  
31 and (e) 20-year-old bauxite residue (R5). Each zone was approximately 1500 m<sup>2</sup>. Within the zones,  
32 natural colonization only occurred in R5. For each zone, five random points, taken within 100 m x 100  
33 m, were designated as the replicates. For each sampling point, the residues were sampled with an auger  
34 to a depth of 20 cm. The samples were then stored in polyethylene bags, returned to the laboratory, air  
35 dried at room temperature for two weeks and then subsequently passed through a 2 mm sieve prior to  
36 analysis.

### 37 *Physical and chemical analysis*

38 Mechanical composition of residue samples were analyzed using a Malvern Mastersizer 2000  
39 (Malvern Instruments Ltd., UK) (Santini and Fey, 2013). pH and electrical conductivity (EC) of residue  
40 samples were determined in 1:5 solid/solution extracts. Exchangeable Ca<sup>2+</sup>, Mg<sup>2+</sup>, K<sup>+</sup>, and Na<sup>+</sup> were  
41 extracted with 1 M ammonium acetate and analyzed by ICP-AES (Jones *et al.*, 2011). Exchangeable  
42

sodium percentage (ESP) was calculated as the percentage of exchangeable Na<sup>+</sup> in the total exchangeable cations. The total contents of Ca, Mg, K and Na in bauxite residue were determined after microwave digestion using HF, HCl and HNO<sub>3</sub> and analysed by ICP-AES (Jones *et al.*, 2011). Total organic carbon was measured by the low-temperature external-heat potassium dichromate oxidation colorimetric method (Zhu *et al.*, 2016a). Chemical phases of residue samples were determined by X-ray powder diffraction (XRD) on a Bruker D8 discover 2500. XRD patterns were collected from 10° to 80° at a step size of 0.04° 2θ with a scan rate of 1° 2θ per minute and analysed using PANalytical analysis package (Zhu *et al.*, 2017).

#### *Microaggregate Stability Analysis*

Laser sizing (for the <0.25 mm fraction) was used to determine particle size distribution of residue microaggregates (Santini & Fey, 2013). In this method, 10 g of air-dried residue samples were placed in a 0.25 mm sieve. The residue samples were then immersed in distilled water and oscillated for 24 h using an end-over-end shaker with a rate of 200 cycles per minute. Particle size distribution of the <0.25 mm aggregates was determined using a Malvern Mastersizer 2000. In order to observe residue microaggregate distribution characterization under natural regeneration, micro-morphological studies of the residue microaggregates from R1 and R5 were examined using a FET Quanta-200 scanning electron microscope (SEM), equipped with energy dispersive X-ray spectroscopy. The specimen was sputter coated with a layer of gold prior to examination (Zhu *et al.*, 2016b).

Water-dispersible clay (WDC) and water-dispersible silt (WDSI) were determined as the proportion of clay and silt in suspension in the distilled water. Clay dispersion ratio (CDR) and aggregated silt+clay indices (ASC) were selected as the two indicators to measure microaggregate stability of bauxite residue. Clay dispersion ratio (CDR) was determined as the following equation (Cammeraat & Imeson, 1998):

$$CDR(\%) = \frac{\%clay + \%silt(\text{water dispersed})}{\%clay + \%silt(\text{calgon dispersed})} \times 100 \quad (1)$$

This is defined as the percentage ratio of clay+silt (<0.02 mm) obtained from both distilled water and sodium hexametaphosphate (calgon) dispersed residue samples. The value of ASC was negatively correlated to aggregate stability (Mbagwu & Auerswald, 1999).

Aggregated silt and clay (ASC) was calculated using the following equation:

$$ASC(\%) = (\%clay + \%silt)(\text{calgon dispersed}) - (\%clay + \%silt)(\text{water dispersed}) \quad (2)$$

A higher ASC value indicates greater microaggregate stability (Monreal *et al.*, 1995).

#### *Calculation of Single-fractal Dimension (D)*

The power-law relationship between either number-diameter, mass-diameter or bulk density-diameter of soil aggregates are always used to determine the fractal dimension of soil aggregates. Here, according to Tyler & Wheatcraft (1989), mass-diameter of residue aggregates was selected to calculate the fractal dimension of microaggregates, designated as D, as follows:

$$D = 3 - \lg(W_i/W_o) / \lg(\bar{d}_i/\bar{d}_{max}) \quad (3)$$

where D is the mass fractal dimension; W<sub>i</sub> is the cumulative mass of the <d<sub>i</sub> residue aggregates; W<sub>o</sub> is the total mass of the residue aggregates;  $\bar{d}_i$  is the mean diameter of aggregates in adjacent particles

1 and  $\bar{d}_{\max}$  is the mean diameter of the largest aggregates.

### 3 Calculation of Multi-fractal Parameters

4 In this study, the measurement interval of the laser particle size analyzer ( $I=[0.01 \mu\text{m}, 250 \mu\text{m}]$ )  
 5 was considered as the residue microaggregate size volume percentages obtained from the previous  
 6 results. The microaggregate size interval is divided into 74 subintervals  $I_i=[\varphi_i, \varphi_{i+1}]$ ,  $i=1, 2, \dots, 74$ .  
 7 Based on standard microaggregate-size division methods,  $\log(\varphi_{i+1}/\varphi_i)$  is the constant following the  
 8 measurement interval of  $I=[0.01, 250]$ . In order to build a new measurement of the multi-fractal  
 9 method,  $\psi_i=\log(\varphi_i/\varphi_1)$  ( $i=1, 2, \dots, 74$ ) was created to form a new dimensionless interval of  $J=[0, 4.40]$ ,  
 10 which had 74 subintervals of equal length,  $J_i=[\psi_i, \psi_{i+1}]$  ( $i=1, 2, \dots, 74$ ). In the interval  $J$ ,  $\varepsilon$  was defined  
 11 as  $2^k$  same size subintervals,  $\varepsilon=4.4 \times 2^{-k}$ . The value of  $k$  ranged from 1 to 6 to make sure that every  
 12 subinterval contained at least one measured value (Peng *et al.*, 2014). Thus, the multi-fractal  
 13 parameters including capacity dimension ( $D_0$ ), information dimension ( $D_1$ ), correlation dimension ( $D_2$ )  
 14 and information dimension/capacity dimension ( $D_1/D_0$ ) were calculated as the following equations  
 15 (Ahmadi *et al.*, 2011):

$$16 \quad D(q) \approx \lim_{\varepsilon \rightarrow 0} \frac{1}{q-1} \times \frac{\log \left[ \sum_{i=1}^{n(\varepsilon)} u_i(\varepsilon)^q \right]}{\log \varepsilon} \quad q \neq 1 \quad (4)$$

$$17 \quad D_1 \approx \frac{\sum_{i=1}^{n(\varepsilon)} u_i(\varepsilon) \log u_i(\varepsilon)}{\log \varepsilon} \quad q = 1 \quad (5)$$

18 The value of  $q$  varied between  $-10$  and  $10$  with a step size of  $1$ . The multi-fractal spectrum of the  
 19 residue microaggregate size distribution were determined by  $D(q)$ .  $D_0$  indicated the span of the residue  
 20 microaggregate size distribution and the larger  $D_0$  value representing a wider range;  $D_1$  indicated the  
 21 irregular degree of the residue microaggregate size distribution and the higher  $D_1$  representing a higher  
 22 level of dispersion in microaggregate size distribution;  $D_1/D_0$  can measure the degree of heterogeneity  
 23 of microaggregate size distribution (Peng *et al.*, 2014).

### 24 Data Analysis

25 All analyses were performed in quintuplicate. The data were statistically treated with Microsoft  
 26 Excel 2003, SPSS version 19.0 and Origin 8.0. A two-way ANOVA, followed by Tukey's post hoc test  
 27 was used to determine the interaction between fragment size range and residue disposal ages. Chemical  
 28 properties of bauxite residue samples with different chronosequences were individually determined  
 29 using one-way ANOVA followed by Tukey's post hoc tests. In the case of no homogeneity, Dunnett's  
 30 T3 test was performed. Bivariate correlation analyses were used to determine the relationships between  
 31 fractal parameters and residue microaggregate size distribution. All figures were constructed using  
 32 Origin 8.0.

## 34 RESULTS AND DISCUSSION

### 35 Particle Size Distribution of Residue Microaggregates

36 The particle size distribution of microaggregates is shown in Table I. The main fraction,  $<0.02$   
 37 mm aggregates, accounted for more than 55% of the total microaggregate weight in newly stacked

1 residue (R1). Microaggregate fractions in the fresh residue (R1) decreased in the following order,  
2 250-50  $\mu\text{m}$  > 10-5  $\mu\text{m}$  > 20-10  $\mu\text{m}$  > 5-2  $\mu\text{m}$  > 50-20  $\mu\text{m}$  > <2  $\mu\text{m}$ . With increasing disposal age, the  
3 aggregate fraction 250-50  $\mu\text{m}$ , increased significantly from 27.4% to 40.3%, whilst the clay-size  
4 aggregate fraction decreased gradually. Microaggregate fractions which had been disposed for 20 years  
5 (R5) decreased in the following order, 250-50  $\mu\text{m}$  > 50-20  $\mu\text{m}$  > 20-10  $\mu\text{m}$  > 10-5  $\mu\text{m}$  > 5-2  $\mu\text{m}$  > <2  
6  $\mu\text{m}$ . With increasing disposal age, the fine particles of bauxite residue became coarser and the silt+clay  
7 size (<0.02 mm) aggregate fraction effectively decreased.

8 SEM images of residue microaggregates from R1 and R5 are shown in Fig. 1. The residue  
9 microaggregate in R1 contained numerous amorphous substances and fine particles. Compared to R1,  
10 the microaggregate in R5 consisted of a great number of larger aggregates. With increasing disposal  
11 age, the residue microaggregates had a denser structure and the particles were distributed uniformly.  
12 Total sodium content decreased from 9.27% to 1.07%, whilst total calcium content increased from  
13 13.89% to 27.88% (Fig. 1), which suggested that natural weathering processes decreased sodium but  
14 increased calcium content. The result from Fig. 1 was consistent with the variation trend of calcium and  
15 sodium content in Table II.

16 Natural weathering processes may have a positive effect on particle aggregation; the finer particles  
17 aggregating to form larger particles (Santini & Fey, 2013). Climate affects soil aggregation through  
18 alterations in temperature and moisture regimes and wet-dry and freeze-thaw cycles, which can  
19 re-orientate particles and improve aggregation (Singer *et al.*, 1992). Weathering alters materials, which  
20 are translocated within the soil through leaching, eluviation, and illuviation resulting in horizonation  
21 (Garcia-Franco *et al.*, 2015; Zhou *et al.*, 2017). Plant roots and their rhizospheres have positive effects  
22 on soil aggregation. Roots realign soil particles and release exudates, which result in physical, chemical  
23 and biological alterations that influence aggregation (Rillig *et al.*, 2001). Courtney *et al.* (2009)  
24 revealed that gypsum and spent mushroom compost application may decrease microaggregate  
25 breakdown and hence the dominance of less erodible aggregates. Courtney *et al.* (2013) found that  
26 addition of gypsum and compost produced a significant decrease in clay- and silt-size particles, whilst  
27 mean particle size (<53  $\mu\text{m}$ ) was lowest in the unamended residues. Natural weathering processes  
28 may ameliorate physical and chemical properties of bauxite residue, which lead to clay-size particles  
29 flocculating and the formation of more stable aggregates.

### 31 *Microaggregate Stability of the Residues*

32 Colloidal stability indices of the residues are presented in Fig. 2. Water-dispersible clay (WDC)  
33 which was used to estimate microaggregate instability, ranged from 0.64% to 2.04%, whilst  
34 water-dispersible silt (WDSI), also used to estimate instability, ranged from 30.72% to 58.39%. A  
35 combination of WDC and WDSI gave values of between 31.4% and 58.4%. Clay dispersion ratio  
36 (CDR) of the residues ranged from 7.7% to 22.5%, and aggregated silt and clay (ASC) ranged from  
37 15.3% to 19%. Following natural disposal processes, WDC, WDSI, CDR and ASC of the residues  
38 stacked for 20 years (R5) increased by 218.75%, 90.07%, 192.21% and 24.18%, respectively.

39 The WDC, WDSI, CDR and ASC may be used to estimate the rate of soil dispersibility. A high  
40 WDC and dispersion indices have negative implications for the soil environment in terms of water and  
41 wind erosion (Virto *et al.*, 2008). With increasing disposal age, WDC, WDSI and CDR significantly  
42 decreased, whilst ASC increased indicating that natural stacking processes may improve  
43 microaggregate stability; the finer particles may combine together to form larger particles due to  
44 binding agents related to physical and chemical properties of the residues (Zhu *et al.*, 2016c). Plant

1 growth and root penetration may also have positive effects on particle aggregation and stability as the  
2 residues stacked for 20 years (R5) had improved colloidal stability compared to the other locations.

3 Stability of microaggregates, as opposed to its dispersion, is a very important soil property that  
4 regulates soil degradation. Several major binding agents including clay content, organic carbon and  
5 electrolytes had significant effects on microaggregate stability. Clay content, organic carbon, pH and  
6 exchangeable cations were selected to identify correlational relationships with colloidal stability  
7 indices. The selected physical and chemical properties were determined in a previous study (Zhu *et al.*,  
8 2016c). Silt and clay contents ranged from 48.8%-23.8% and 5.9%-1.5% respectively. With increasing  
9 disposal age, pH and EC were significantly reduced. Total organic carbon (TOC) content ranged from  
10 5.7-10.8 g/kg. Exchangeable Ca and Na varied regularly but with opposite trends (Table II).

11 As natural weathering processes had a significant effect on bauxite residue mineral chemistry, the  
12 residue samples, including R1 and R5, were selected to investigate the variation in chemical phases  
13 (Figure 3). Slaked lime addition by the Bayer process resulted in the formation of calcium minerals  
14 including calcite ( $\text{CaCO}_3$ ), hydrogarnet ( $\text{Ca}_3\text{Al}_2(\text{SiO}_4)_x(\text{OH})_{12-4x}$ ), and tri-calcium aluminate  
15 ( $\text{Ca}_3\text{Al}_2(\text{OH})_{12}$ ). Other alkaline minerals in the residue included sodalite ( $\text{Na}_6\text{Al}_6\text{Si}_6\text{O}_{24} \cdot [2\text{NaOH}$  or  
16  $\text{Na}_2\text{CO}_3]$ ) and cancrinite ( $\text{Na}_6\text{Al}_6\text{Si}_6\text{O}_{24} \cdot 2\text{CaCO}_3$ ). Following natural processes, sodalite, hydrogarnet  
17 and calcite decreased, as did pH. Reduction of pH resulted in the precipitation of  $\text{Ca}(\text{OH})_2$  but also  
18 leaching of NaOH by exchange reactions as the charged colloid (such as  $\text{Al}(\text{OH})_6^{3-}$ ) was a regulator for  
19 cation exchange (Kong *et al.*, 2017). As a result, exchangeable sodium percentage decreased. In  
20 addition, sodium ions could not be coordinated with negatively charged surfaces which led to the  
21 formation of alkaline dust, reduction on erosion resistance and a poor physical structure (Zhu *et al.*,  
22 2016d).

23 Linear regression analysis showed that the value of clay dispersion ratio (CDR) was positively  
24 correlated to clay content, pH and exchangeable  $\text{Na}^+$  content ( $r=0.898, 0.943, \text{ and } 0.826$  respectively;  
25  $P<0.01$ ), but negatively correlated to exchangeable  $\text{Ca}^{2+}$  and total organic carbon content ( $r = -0.972$   
26 and  $-0.936, P<0.01$ ) (Fig. 4). The value of aggregated silt and clay (ASC) was negatively correlated to  
27 clay content, pH and exchangeable  $\text{Na}^+$  content ( $r = -0.903, -0.927, \text{ and } -0.865, \text{ respectively; } P<0.01$ ),  
28 and positively correlated to exchangeable  $\text{Ca}^{2+}$  and total organic carbon content ( $r = 0.948 \text{ and } 0.932$   
29 respectively,  $P<0.01$ ) (Fig. 5). This indicated that high exchangeable  $\text{Ca}^{2+}$  content and low  
30 exchangeable  $\text{Na}^+$  stimulated microaggregate flocculation, whilst the decrease in pH and the  
31 accumulation of organic carbon may have improved microaggregate stability. Courtney *et al.* (2009)  
32 established a field scale investigation to promote vegetation cover on bauxite residue, and found that  
33 spent mushroom compost and gypsum amendments decreased pH and ESP which positively impacted  
34 on microaggregate stability. Addition of Ca had a positive effect on flocculating clay particles, reducing  
35 mechanical dispersion and lowering exchangeable  $\text{Na}^+$  content, thereby stabilizing microaggregates  
36 (Harris & Rengasamy, 2004). Pojasok & Kay (1990) reported that increasing organic carbon content  
37 stimulated particle aggregation.

### 39 *Single-fractal Features of Residue Microaggregates*

40 The fractal dimension of soil microaggregates may reflect the geometry parameters of soil  
41 aggregate structure, with higher clay content indicating the higher value of the fractal dimension  
42 (D). Under natural weathering processes, residue fractal dimension (D) was significantly affected  
43 (Fig. 6). The single-fractal dimension ranged from 2.2 to 2.4 and with increasing disposal age,  
44 microaggregate fractal dimension (D) decreased. R1 had a low proportion of 250-20  $\mu\text{m}$  size

1 aggregates and a high proportion of 10-2  $\mu\text{m}$  aggregates, which resulted in a high fractal  
2 dimension value. Under natural soil forming processes, fine particle aggregation led to the  
3 decrease in single-fractal dimensions. Certainly, single-fractal dimension (D) was positively  
4 correlated with the proportion of 10-5  $\mu\text{m}$ , 5-2  $\mu\text{m}$ , and <2  $\mu\text{m}$  sized microaggregates ( $r=0.859$ ,  
5  $0.977$ ,  $0.991$  respectively,  $P<0.01$ ), but negatively correlated with the proportion of 250-50  $\mu\text{m}$   
6 and 50-20  $\mu\text{m}$  sized microaggregates ( $r=-0.876$  and  $0.761$  respectively,  $P<0.01$ ). There was no  
7 significant correlation between single-fractal dimension and the proportion of 20-10  $\mu\text{m}$  sized  
8 microaggregates. Single-fractal dimension may be regarded as an important indicator to reflect  
9 aggregate structure of bauxite residue.

10 High alkalinity and salinity resulted in poor aggregate structure of the residue and clearly  
11 affected revegetation on disposal areas (Jones *et al.*, 2011). Zhu *et al.* (2016d) found that natural  
12 vegetation encroachment ameliorated residue physicochemical properties and stimulated  
13 aggregate stability. Courtney *et al.* (2013) investigated the physical condition of revegetated  
14 residue and found that gypsum and organic carbon decreased pH and ESP, which enhanced the  
15 proportion of water-stable aggregates, which supported plant growth. The related relationships  
16 between single-fractal dimension and pH, EC, and ESP are displayed in Fig. 7. The single-fractal  
17 dimension was positively correlated with pH, ESP, exchangeable  $\text{Na}^+$  content and EC ( $r=0.935$ ,  
18  $0.984$ ,  $0.859$  and  $0.912$  respectively,  $P<0.01$ ), but negatively correlated with exchangeable  $\text{Ca}^{2+}$   
19 content ( $r=-0.968$ ,  $P<0.01$ ). It indicated that the single-fractal dimension of residue  
20 microaggregates may reflect the related physical and chemical properties of bauxite residue.

#### 21 22 *Multi-fractal Dimension of Residue Microaggregates*

23 Multi-fractal spectrums of residue microaggregate size distributions between -10 and 10 at 1.0 lag  
24 increments for different disposal ages are presented in Fig. 8. The multi-fractal spectrums show a  
25 typical anti-S-decreasing function. The information entropy ( $D_{(q)}$ ) of residue microaggregates  
26 decreased with increasing disposal age. Furthermore, in each residue sample,  $D_0>D_1$  always existed,  
27 meaning that microaggregate size distribution with different disposal ages were not homogeneous or  
28 monofractal. Therefore, multi-fractal dimension analysis was essential.

29 The value of  $D_0$ ,  $D_1$  and  $D_1/D_0$  decreased with increasing disposal age (Table III). In residues  
30 which had been stacked for 20 years (R5), these values were nearly the lowest ( $D_0$ ,  $D_1$  and  $D_1/D_0$   
31 values of  $0.942$ ,  $0.853$  and  $0.906$ , respectively), whilst for R1, these values were the highest ( $D_0$ ,  $D_1$   
32 and  $D_1/D_0$  values of  $0.968$ ,  $0.891$  and  $0.920$ , respectively). Analysis of variance in the different residues  
33 showed that the multi-fractal parameters of R1 and R2 were significantly different ( $P<0.05$ ).

34 The larger  $D_0$  means a wider range of microaggregate size distributions. Nevertheless, the  
35 calculation of  $D_0$  is based on the assumption that particle size distribution was homogeneous. The  
36 value of  $D_1/D_0$  may make a quantitative description of the heterogeneous degree of soil particle  
37 size distribution. Miranda *et al.* (2006) pointed out that if the value of  $D_1/D_0$  was closer to 1 this  
38 specified that the particle size distribution was more concentrated. The value of  $D_1/D_0$  in the  
39 residue microaggregates ranged from  $0.896$  to  $0.920$ . With increasing disposal age, the value of  
40  $D_1/D_0$  decreased indicating that natural weathering processes decreased the concentration of  
41 microaggregate distribution. Natural processes accumulate organic carbon over time and this may  
42 have ameliorated the high alkalinity and salinity in the residue, stimulating fine particle  
43 aggregation. A significant increase in the proportion of 250-50  $\mu\text{m}$  residue microaggregates resulted  
44 in homogeneity with increasing disposal age.

1 Bivariate correlation analysis between multi-fractal parameters and residue microaggregate  
 2 distribution showed that the value of  $D_0$ ,  $D_1$ ,  $D_1/D_0$  was significantly correlated with the  
 3 proportion of  $<2 \mu\text{m}$  microaggregates ( $r=0.915$ ,  $0.786$  and  $0.523$  respectively,  $P<0.05$ ). In addition, the  
 4 value of  $D_1$  was positively correlated with the proportion of  $10\text{-}5 \mu\text{m}$  and  $5\text{-}2 \mu\text{m}$  microaggregates  
 5 ( $r=0.912$  and  $0.671$  respectively,  $P<0.05$ ). According to correlation analysis between multi-fractal  
 6 parameters,  $D_1$  was positively correlated with  $D_0$  and  $D_1/D_0$  ( $r=0.933$  and  $0.917$  respectively,  $P<0.05$ ).  
 7 Multi-fractal parameters of residue microaggregate size distribution were mainly affected by the  
 8 proportion of  $<10 \mu\text{m}$  microaggregates, especially the silt-sized ( $<2 \mu\text{m}$ ) microaggregates.

9 Multi-fractal dimension of residue microaggregates was closely linked with related  
 10 physicochemical properties. Residues which had been stacked for 20 years had a low pH and EC which  
 11 led to a low multi-fractal dimension, whilst the newly stacked residue had a high pH and EC and a high  
 12 multi-fractal dimension. This suggested that the multi-fractal parameters of microaggregates may  
 13 reflect physical and chemical properties and may be used as an effective indicator to characterize  
 14 alkalinity and salinity of bauxite residue. According to multiple linear models between multi-fractal  
 15 parameters and the related properties of bauxite residue, the following equations were obtained:

$$16 \quad D_0=1.01926+0.0018x_1-0.0079x_2-0.00376x_3+2.06\times 10^{-4}x_4 \quad (6)$$

$$17 \quad D_1=0.91468+0.00325x_1-0.00476x_2+1.9669\times 10^{-4}x_3-3.91826\times 10^{-4}x_4 \quad (7)$$

$$18 \quad D_1/D_0=0.88054+0.00498x_1-0.0019x_2+0.00231x_3+6.58484\times 10^{-5}x_4 \quad (8)$$

19 where  $x_1$  is the content of TOC,  $x_2$  is the value of pH,  $x_3$  is the value of EC, and  $x_4$  is the value of ESP.  
 20 This demonstrated that organic carbon content and pH were the main properties influencing the values  
 21 of multi-fractal dimension of residue microaggregates.

### 22 *Relationship between Microaggregate Stability and Fractal Parameters*

23 Microaggregate stability is usually used to estimate or predict soil erosion and surface runoff  
 24 (Wang *et al.*, 2016). Bauxite residue has poor physical structure to resist water erosion and support  
 25 revegetation. Zhu *et al.* (2016d) discovered that following natural weathering processes, the erodibility  
 26 factor of the residue decreased indicating improved resistance to erosion. Correlation analysis showed  
 27 that single- and multi-fractal parameters were significantly correlated to microaggregate stability  
 28 indicating that fractal parameters of residue microaggregate distribution may reflect microaggregate  
 29 stability. The fractal dimension of microaggregate size distribution may exhibit variation in  
 30 microaggregate size distribution. The high value fractal parameters indicated a dense physical structure  
 31 and poor erosion resistance.

32 Tang *et al.* (2013) revealed a significant negative relationship between fractal dimension and soil  
 33 microaggregate content ( $<0.25 \text{ mm}$ ) in karst rocky desertification areas and suggested that fractal  
 34 dimension could be used as a reliable indicator of soil quality. A small fractal dimension value for  
 35 granular structure indicated a stable soil structure. Ahmadi *et al.* (2011) found that both number- and  
 36 mass-based fragmentation fractal dimension may describe the aggregate size distribution and estimate  
 37 splash and inter-rill soil erosion. In our study, the single-fractal dimension ( $D$ ) of microaggregate  
 38 distribution was negatively correlated with ASC ( $r=-0.977$ ,  $P<0.01$ ), whilst positively correlated with  
 39 CDR ( $r=0.995$ ,  $P<0.01$ ). The value of  $D_0$  and  $D_1$  showed a significant correlation with ASC ( $r=-0.823$   
 40 and  $-0.739$  respectively,  $P<0.01$ ) and CDR ( $r=0.822$  and  $0.709$  respectively,  $P<0.01$ ), whilst  $D_0/D_1$  had

1 little significant difference with ASC and CDR (Table IV). With increasing disposal age, aggregation  
2 of fine particles resulted in a lower value of fractal dimension and a more stable aggregated structure.  
3 This suggests that fractal dimension may be useful to characterize microaggregate stability. Compared  
4 to multi-fractal parameters, single-fractal dimension (D) was more significantly correlated with  
5 microaggregate stability in bauxite residue.

## 6 CONCLUSIONS

7 Microaggregate stability, an important physical indicator, is required to sustain a stable physical  
8 structure. This study has clearly demonstrated that natural weathering processes significantly affect  
9 particle size distribution of residue microaggregates. With increasing disposal age, the proportion of  
10 silt- and clay-sized microaggregates significantly decreased. Clay dispersion ratio (CDR) decreased  
11 from 22.5% to 7.7%, and aggregated silt and clay (ASC) increased from 15.3% to 19% indicating that  
12 natural weathering processes enhanced microaggregate stability. Clay content, organic carbon,  
13 exchangeable bases and pH were significantly correlated with ASC and CDR which indicated that  
14 organic carbon and exchangeable cations had significant effects on microaggregate stability. The value  
15 for single-fractal dimension (D) varied from 2.2 to 2.4. With increasing disposal age, both single-fractal  
16 dimension (D) and multi-fractal parameters ( $D_0$ ,  $D_1$  and  $D_1/D_0$ ) decreased, revealing that natural  
17 weathering process promoted aggregation of microaggregates. Correlation analyses demonstrated that  
18 fractal parameters were significantly correlated with microaggregate stability and physicochemical  
19 properties, indicating that fractal parameters may be used to characterize residue physical structure and  
20 related properties. This study may help to provide an improved understanding of physical  
21 microstructures, appropriate indicators to use when evaluating microaggregate stability, and a scientific  
22 basis for the revegetation of bauxite residue disposal areas.

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